

1 **Shifts in marine invertebrate bacterial assemblages associated with tissue necrosis**
2 **during a heatwave**

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25 **Abstract**

26 Marine heatwaves (MHWs) are periods of extremely high seawater temperature that
27 affect marine ecosystems in several ways. Anthozoans (corals and gorgonians) and
28 Porifera (sponges) are usually among the taxa most affected by MHWs. Both are
29 holobiont entities that form complex interactions with a wide range of microbes, which
30 are an essential part of these organisms and play key roles in their health status. Here, we
31 determine microbial community changes suffered in two corals (*Cladocora caespitosa*
32 and *Oculina patagonica*), one gorgonian (*Leptogorgia sarmentosa*), and one sponge
33 (*Sarcotragus fasciculatus*) during the 2015 MHW. The microbial communities were
34 different among hosts and displayed shifts related to host health status, with a higher
35 abundance in necrosed tissues of *Ruegeria* species or of potential pathogens like *Vibrio*.
36 We also carry out a meta-analysis using 93 publicly accessible 16S rRNA gene libraries
37 from *O. patagonica*, *C. caespitosa* and *L. sarmentosa* to establish a Mediterranean core
38 microbiome in these species. We have identified one *Ruegeria* OTU that maintained a
39 stable and consistent association with these species, which was also related with tissue
40 necrosis in their hosts. Therefore, *Ruegeria* sp. could play an important and still
41 underexplored role in the health status of its hosts.

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46 **Introduction**

47 Marine heatwaves (MHWs) are periods of extremely high seawater temperature that
48 persist for days to months and can extend up to thousands of kilometers (Frölicher and
49 Laufkötter, 2018). Some of the recently observed marine heatwaves revealed the high
50 vulnerability of marine ecosystems, which can be affected in several ways, such as by
51 decreasing productivity, altering food web dynamics, shifting species distribution, and
52 reducing abundance (Hughes et al., 2003; Hoegh-Guldberg and Bruno, 2010). MHWs,
53 which will probably intensify with anthropogenic climate change (Frölicher and
54 Laufkötter, 2018), are related to mass mortality events and disease outbreaks in marine
55 species that severely threaten the structure and functioning of ecosystems and disrupt the
56 provision of ecological goods and services in coming decades (Smale et al., 2019). The
57 most recently observed marine heatwave with global ecological implications was
58 recorded in 2015/16, when unusually high ocean temperatures associated with one of the
59 strongest El Niño events on record triggered unprecedented coral bleaching and marine
60 invertebrate mortality worldwide (Rubio-Portillo et al., 2016a; Ampou et al., 2017; Oliver
61 et al., 2017; Turicca et al., 2018).

62 Anthozoa (Scleractinians and Octocorals) and Porifera are important members of the
63 benthic community. These taxa provide structural complexity to ecosystems and thereby
64 refuge and habitats to other fauna and are the taxa most affected by MHWs (Cerrano et
65 al., 2000; Garrabou et al., 2009). Like all multicellular organisms, marine benthic
66 invertebrates (encompassing Anthozoa and Porifera) are holobiont entities, forming
67 complex interactions with a wide range of microbes, including dinoflagellates, fungi,
68 bacteria, archaea, and viruses (Knowlton and Rohwer, 2003). These microbial symbionts
69 play active roles in holobiont health (e.g., nutrient supply and protection against

70 pathogens) as well as the adaptive response of the host to environmental changes
71 (reviewed in Bourne et al., 2016 and Pita et al., 2018).

72 Changes in the environment may severely disturb host-microbe interactions and thus lead
73 to dysbiosis (microbial imbalance on or inside the host) and/or disease development
74 (Harvell et al., 2007; Miller and Richardson, 2014; Sweet et al., 2015). Therefore, the
75 evaluation of the shifts in microbiota as a result of MHWs may be employed as “early”
76 bio-indicators of both environmental changes and host disease. However, few studies
77 have investigated the microbiota of marine invertebrates other than corals during
78 warming events. Microbial community association with marine invertebrates is dynamic
79 and includes a ubiquitous core microbiome, which is defined as stable and consistent
80 components across complex microbial assemblages from similar habitats (see review by
81 Sweet and Bulling, 2017). These core members play key roles due to their ability to
82 maintain microbial associations’ stability under environmental changes through
83 competition for nutrients and/or space with invasive microbes, as well as by production
84 of antibiotics (Ritchie, 2006; Krediet et al., 2013). Along with core members, there is a
85 second associated microbial fraction that is more influenced by the local environmental
86 conditions and a third highly variable component dependent on the processes occurring
87 at the spatial and temporal scales (Reveillaud et al., 2014; Ainswoth et al., 2015;
88 Hernandez-Agreda et al., 2018). Given the likely critical contribution of microbes to
89 invertebrate holobiont adaptation to environmental changes, shifts in marine
90 invertebrates’ microbial assemblages could be ideal indicators for host heat stress.

91 In the last 20 years, Mediterranean marine invertebrates have suffered an increase of
92 disease outbreaks due to warming events (Cerrano et al., 2000; Garrabou et al., 2009;
93 Stabili et al., 2012; Jiménez et al., 2016). Specifically, during the 2015 MHW in the
94 Marine Protected Area of Tabarca, more than 40% of the population of the sponge

95 *Sarcotragus fascicualatus*, the corals *Cladocora caespitosa* and *Oculina patagonica*, as
96 well as the gorgonian *Leptogorgia sarmentosa* showed tissue necrosis signs as a
97 consequence of the increase of seawater temperature (Rubio-Portillo et al., 2016a). Thus,
98 the main goal of this study was to assess the effect of marine heat waves on microbial
99 assemblages associated to those species to understand the influence of global warming
100 conditions on these associations and ultimately on the health of marine invertebrates. To
101 achieve this goal, we used Next Generation Sequencing to characterize, by means of 16S
102 rRNA gene metabarcoding, a total of 24 marine invertebrate tissue samples from
103 apparently healthy and necrosed colonies. We have identified potential microbial bio-
104 indicators of marine invertebrate diseases, such as the increase of *Ruegeria* and *Vibrio*
105 genera and a decrease of putative symbionts like *Pseudovibrio* or *Endozoicomonas* in
106 necrosed tissues of corals and gorgonian, respectively. Moreover, a meta-analysis using
107 93 publicly accessible 16S rRNA gene libraries from *O. patagonica*, *C. caespitosa* and *L.*
108 *sarmentosa* was carried out to establish a Mediterranean core microbiome in these
109 species. We determined a stable and consistent association between a *Ruegeria* OTU and
110 geographically and phylogenetic distinct Mediterranean Anthozoans, which could play
111 an important role in the host health status. Further, our results suggest that the
112 composition of the core microbiome depends on the geographical area considered in the
113 analysis, confirming the existence of a local core microbiome that depends on the
114 surrounding environment.

115 **Material and methods**

116 **Sample collection**

117 Water and invertebrate samples were collected on 28 September 2015 in two sampling
118 locations in the Marine Protected Area of Tabarca. The gorgonian *L. sarmentosa* was

119 collected at 25m depth (38°09'35'' N, 00°27'55''E), while the coral *O. patagonica* and
120 the sponge *S. fasciculatus* were collected at 5m depth (38°09'59'' N, 00°28'56''E, Spain).
121 For each of the three sampled marine invertebrates, a total of six (3 healthy and 3
122 necrosed) independent specimens were taken. In addition, two water samples were taken
123 from each sampling location (Table 1). The health status of the invertebrates and the
124 environmental parameters are described in Rubio-Portillo et al., (2016a).

125 All samples were taken during the heat wave recorded in September 2015. During this
126 MHW, water temperature was 2 °C higher compared with the preceding 9 years, and
127 persisted for approximately 6 weeks, reaching a maximum of 28.23°C (Rubio-Portillo et
128 al., 2016a). Marine invertebrate samples were removed by SCUBA diving using a
129 hammer and chisel and placed in plastic bags under water. Two water samples were taken
130 from each depth using sterilized bottles. All samples were transported to the laboratory
131 in a cooler within the next 2 hours. In the lab, marine invertebrate samples were gently
132 washed three times with 50 ml of sterile filtered seawater (SFSW) to remove non-
133 associated microbes and approximately 2 g (wet weight) of each sample was crushed with
134 5 ml SFSW using a mortar and they were allowed to settle for 15min and the supernatant
135 (that is, crushed tissue) was removed and kept at -80 °C for further analyses.

136 **DNA extraction and polymerase chain reaction amplification of 16S rRNA genes**

137 DNA was extracted from crushed tissue using the UltraClean Soil DNA Kit (MoBio;
138 Carlsbad, CA) following the manufacturer's instructions for maximum yield. DNA from
139 water samples was extracted using the DNeasy blood and tissue kit (Qiagen, Valencia,
140 CA). The extracted genomic DNA was used for PCR amplifications of the V3-V4 region
141 of the 16S rRNA gene by using the following universal primers: Pro341F
142 (CCTACGGGNBGCASCAG) (Takahashi et al., 2013) and Bact805R

143 (GACTACHVGGGTATCTAATCC) (Herlemann et al., 2011). Each PCR mixture
144 contained 5 µl of 10x PCR reaction buffer (Invitrogen), 1.5 µl of 50 mM MgCl₂, 1 µl 10
145 mM dNTP mixture, 1 µl of 100 µM of each primer, 1 unit of Taq polymerase, 3 µl of
146 BSA (New England BioLabs), sterile MilliQ water up to 50 µl and 10 ng of DNA.
147 Negative controls (with no template DNA) were included to assess potential
148 contamination of reagents. The amplification products were purified with the GeneJET
149 PCR purification kit (Fermentas, EU), quantified using the Qubit Kit (Invitrogen), and
150 the quality (integrity and presence of a unique band) was confirmed by 1% agarose gel
151 electrophoresis. Sequencing was performed using Illumina Mi-seq Nextera 2x300 bp
152 paired-end run (at Fundació per al Foment de la Investigació Sanitària i Biomèdica,
153 FISABIO, Valencia).

154 **Illumina high-throughput 16S rRNA gene sequence analysis**

155 Paired-end MiSeq sequences of the 22 samples were deposited in the NCBI Sequence
156 Read Archive (SRA) database. Data from the water samples as well as *O. patagonica*, *L.*
157 *sarmentosa* and *S. fasciculatus* were deposited under BioProject PRJNA615777. For
158 comparative purposes, sequences from the coral *C. caespitosa* (BioProject
159 PRJNA407809) were also included in the analysis. These *C. caespitosa* samples were
160 taken at 5 m depth location in the same sampling campaign than the samples listed in
161 Table 1 and were used for a previous biogeography study (Rubio-Portillo et al., 2018).
162 The QIIME 1.8.0 pipeline (Caporaso et al., 2010) was used for data processing.
163 Operational taxonomic units (OTUs) were defined at the level of 99% similarity, close to
164 the threshold used to distinguish species (98.7% similarity in the whole 16S rRNA gene),
165 (Stackebrandt and Ebers, 2006), followed by taxonomy using UCLUST algorithm
166 (Edgar, 2010) with the SILVA reference database (version 132). OTUs classified as
167 chloroplast or mitochondria were removed from the dataset. Due to the large difference

168 in library size among samples, the OTU table was rarefied to 11,594 reads (the lowest
169 number of the post-assembly and filtered sequences in a sample, Table S1) for
170 comparisons across samples (Weiss et al., 2015).

171 **Analysis of alpha-diversity**

172 Prokaryotic α -diversity was estimated in QIIME prior to deleting singletons and OTUs
173 with less than 0.05% of abundance. Specifically, diversity was characterized using the
174 Shannon diversity index and OTU richness. Differences in alpha diversity index were
175 statistically evaluated using ANOVA analysis in R with the ‘vegan’ package (Oksanen,
176 2011). Prior to ANOVA, homogeneity of variance was confirmed with Cochran’s test
177 (Cochran, 1951) and data was analyzed according to a two-factor model, where the main
178 factors were host (i.e. marine invertebrate species) and health status. If the variances were
179 significantly different at $p = 0.05$, post-hoc analyses were conducted using Student–
180 Newman–Keuls (SNK) multiple comparisons (Underwood, 1997).

181 **Analysis of beta-diversity**

182 Prior to analysis of β -diversity, singletons and OTUs with less than 0.05% of abundance
183 were removed from the dataset. For β -diversity analysis, we used QIIME software and
184 clustering based on the weighted UniFrac (Lozupone and Knight, 2005). To visualize
185 microbiota similarity, we generated principal coordinate analysis (PCoA) plots from the
186 distance matrices. Multivariate analyses were used to compare composition of microbial
187 communities associated with the different marine invertebrate species. Similarity
188 percentage (SIMPER) was used to identify OTUs that could be potentially responsible
189 for these differences.

190 **Core microbiome meta-analysis**

191 In order to identify the core microbiome in the studied area, phylotypes consistently
192 present in 100% of the samples (both healthy and necrosed) from each holobiont were
193 considered. We used a conservative representation of the core microbiome because only
194 six samples were recovered from each marine invertebrate during this study. In addition,
195 to identify cosmopolitan microorganisms associated with benthic Mediterranean
196 invertebrates, the core microbiome of *C. caespitosa*, *O. patagonica* and *L. sarmentosa*
197 across the Mediterranean Sea was analyzed using the recommended cut off at 85% of the
198 samples (Ainswoth et al., 2015; Hernandez-Agreda et al., 2016). For this purpose, a total
199 of 93 16S rRNA libraries were analyzed (12 generated in the present work and 81
200 previously published (Rubio-Portillo et al., 2016b; 2018; van de Water et al., 2017;
201 Bednarz et al., 2019); Table S2). In addition, unique and shared taxa (at the OTU level)
202 among hosts were displayed with the “UpSet” (visualizing intersecting sets) diagram
203 using the “R- bioconductor” package “UpSetR” (Lex et al., 2014).

204 **Results and discussion**

205 To assess the effect of global warming on marine invertebrates, we investigated the
206 differences in the microbiome of healthy and necrosed marine invertebrates during a
207 marine heatwave in order to explore the presence of potential microbial indicators of heat
208 stress. In addition, the core microbiome of each host was also described as well as the
209 presence of cosmopolitan microorganisms associated with benthic Mediterranean
210 Anthozoans.

211 More than 24,000 OTUs were identified in the present study but only 173 OTUs showed
212 a relative abundance over 0.05 % and are discussed here. Invertebrate species hosted on
213 average from 117 to 149 OTUs (149 OTUs for *O. patagonica*, 144 for *L. sarmentosa*, 134
214 OTUs for *C. caespitosa* and 117 for *S. fasciculatus*). Importantly, less than 20% of OTUs

215 identified were host-specific, while about half were shared by at least three of the
216 invertebrates studied here (Figure 1). Particularly, the three Anthozoans are the hosts that
217 shared more OTUs among them (14.45%). Furthermore, 28 OTUs were shared among all
218 the marine invertebrates and seawater samples (Figure 1). Therefore, it seems that the
219 surrounding water has a great influence on the invertebrate microbiome, which is in good
220 agreement with previous studies that showed biogeographical changes in the invertebrate
221 microbiome (Littman et al., 2009; Pantos et al., 2015; Rubio-Portillo et al., 2018). A large
222 proportion of sequences related to the *O. patagonica* pathogen *V. mediterranei*
223 (Kushmaro et al., 1997;1998; Rubio-Portillo et al., 2014) was detected in seawater
224 samples and this OTU was shared by all samples (Table S3). This fact was probably as
225 consequences of the increasing temperature during the MHW and this could compromise
226 benthic invertebrate health. Conversely, since vibrios have been detected in viable but not
227 culturable state in coral tissue during cold seasons (Sharon and Rosenberg, 2010; Rubio-
228 Portillo et al., 2016b), invertebrates could act as a pathogen reservoir, from which they
229 could be dispersed into the surrounding water.

230 The Shannon diversity index ranged from 2 to 5 in the marine invertebrates studied here,
231 consistent with previous studies (Rubio-Portillo et al., 2016b, Thomas et al., 2016; 2018;
232 van de Water et al., 2018a). The two-way ANOVA revealed significant differences
233 among hosts ($F= 23.114$, $p < 0.001$). Post-hoc SNK test showed that these differences
234 were due to the highest diversity values showed by *O. patagonica* compared with the
235 other hosts, which diversity was similar among them (Figure 2A). Similarly, OTU
236 richness was also higher in *O. patagonica* than in the other hosts (Figure 2B; $F= 14.546$,
237 $p < 0.001$). Principal coordinate analysis using weighted UniFrac distances (Lozupone
238 and Knight, 2005) clearly separated samples by hosts, which were also different from
239 seawater samples (Figure 3A and B). Bacterial microbiomes associated with the two

240 zooxanthellate scleractian corals were similar to each other and different from the
241 azooxanthellate gorgonian *L. sarmentosa* microbiome (Fig. 3A and 3B). For instance,
242 *Endozoicomonas* genus, a common coral symbiont (Bourne et al., 2016; Neave et al.,
243 2016), was one of the most abundant genera in the gorgonian *L. sarmentosa* (Fig. 4B and
244 Table 1), in good agreement with previous studies carried out in the Mediterranean Sea
245 (Bayer et al., 2013; Rubio-Portillo et al., 2016b; Rubio-Portillo et al., 2018; Van de Water
246 et al., 2017). However, intriguingly, this genus was absent from the corals studied here.
247 In addition to differences among Anthozoans, differences between the two coral species
248 were also observed. SIMPER analysis showed that *Maritimimonas* was a characteristic
249 genera of *C. caespitosa*, while *Pseudovibrio* genus was significantly enriched in *O.*
250 *patagonica* (Table S4). Likewise, SIMPER analysis revealed that sequences
251 corresponding to uncultured genera of *Acidobacteria* and *Dadabacteria* were sponge-
252 specific (Table S4). Therefore, although surrounding water had a great influence on the
253 invertebrate microbiome, the microbial composition was different for the different hosts
254 and specific symbionts were detected in each host.

255 **Microbiota shifts related to host health status**

256 As shown in figures 2 and 3, although there were no detectable differences in terms of
257 diversity indexes, microbial composition changed depending on health status. Thus, both
258 Shannon index and OTU richness did not show significant changes among healthy and
259 necrosed samples in either host species (Fig. 2A and 2B; $F= 0.796$, $p =0.5141$). However,
260 PERMANOVA analysis ($R^2= 0.651$, $p < 0.005$) as well as principal coordinate analysis
261 using weighted UniFrac distances showed that microbial composition differed depending
262 on health status (Fig. 3A and 3B). SIMPER analysis was carried out in order to detect the
263 OTUs primarily responsible for these differences (Table S5). For the analyzed
264 Anthozoans, this analysis unveiled a common pattern in the necrosed tissues compared

265 with the healthy ones, with a decrease of potential symbionts like *Pseudovibrio*,
266 *Fabibacter* or *Endozoicomonas* genera and an increase of opportunistic and *Vibrio* species,
267 together with the increase of some species whose role is still not clear, like *Ruegeria* spp.
268 (Fig. 4B, Table S6 and Table S4).

269

270 *Pseudovibrio* species, which play a key role in coral health by inhibiting pathogens'
271 growth (Nissimov et al., 2009; Rypien et al., 2010), were more abundant in healthy corals
272 than in necrosed ones. Thus, the *Pseudovibrio*-dominated community changes to a
273 community dominated by potential pathogens in *O. patagonica* necrosed samples. The
274 same pattern was also observed in samples collected in the same studied area in 2011
275 (Rubio-Portillo et al., 2016b). Therefore, this genus appears to be a vital member of the
276 *O. patagonica* holobiont and its abundance could be an indicator of host health.

277 *Pseudovibrio* OTUs detected in the two coral species studied here were different, which
278 suggests that different coral species could select different symbionts in the same
279 environment. In addition to *Pseudovibrio*, *Fabibacter* showed higher abundance in
280 healthy specimens of *C. caespitosa* and it could also play a key role in host health.

281 *Fabibacter* species has been previously reported associated to other coral species
282 (Sunagawa et al., 2009; De castro et al., 2010), but their role remains unknown. For the
283 gorgonian *L. sarmentosa*, tissue damage was associated with a decrease of species
284 commonly associated with gorgonians like *Endozoicomonas* (Figure 4B and Table S4).

285 *Endozoicomonas* spp. are one of the main constituents of octocoral microbial assemblages
286 in the Mediterranean Sea (that can make up to over 96% of the community) and a decrease
287 in its abundance has been correlated to environmental stress (reviewed in van de Water
288 et al., 2018b). This is one of the key findings of this work and highlights the importance
289 of *Pseudovibrio*, and probably *Fabibacter*, together with *Endozoicomonas* genera in

290 Mediterranean Anthozoans. These microbes could serve as potential indicators of
291 compromised health status in Mediterranean corals and gorgonians, respectively.

292 Intriguingly, although the relative abundance of *Vibrio* spp. increased in *O. patagonica*
293 and slightly in *C. caespitosa* necrosed samples (Figure 4B and Table S6), SIMPER
294 analysis did not detect any specific *Vibrio* OTU as primarily responsible for these
295 differences (Table S4). For example, the coral pathogens *Vibrio mediterranei* and *Vibrio*
296 *coralliilyticus* (OTUs 163 and 165, respectively) were detected in necrotic tissues but also
297 in apparently healthy specimens at the same location (Table S5). This result suggests that
298 probably the strains detected in healthy and necrosed samples could be different and not
299 all of them pathogenic. Indeed, *V. mediterranei* strains similar to the type strain AK-1,
300 the causative agent of mass bleaching events in *O. patagonica*, were mainly isolated from
301 the necrosed specimens of *O. patagonica* (Rubio-Portillo et al., 2016). Along with the
302 increase of *Vibrio* species, we have detected a consistent increase of *Ruegeria* sp.
303 SOEmb9 OTU119 in necrosed tissues of all Anthozoans studied here (Table S5).

304 Previous studies have shown that the presence of *Ruegeria* species is correlated with the
305 presence of *Vibrio* pathogens in coral tissues (Rosado et al., 2019), as well as with
306 different signs of disease, such as Black Band Disease in the Caribbean Sea (Sekar et al.,
307 2008), Yellow Band Disease in the Red Sea (Aprill et al., 2013) or White Patch
308 Syndrome in the Indian Ocean (Séré et al., 2013). Furthermore, *Ruegeria* genus,
309 belonging to the *Roseobacter* group, displays high chemotactic attraction towards
310 dimethylsulfoniopropionate (DMSP) (Miller et al., 2004), which is a compound found in
311 heat-stressed zooxanthellate corals (Raina et al., 2013), This behavior could explain the
312 increase of *Ruegeria* sp. in zooxanthellate coral necrotic tissues during this mortality
313 event, probably due to the increase of DMSP as a result of the increase of sea water
314 temperature during the heatwave. However, there are alternative explanations since some

315 studies showed that *Ruegeria* spp. have an important role protecting corals against
316 pathogenic *Vibrio* species by inhibiting their growth (Miura et al., 2018; Rosado et al.,
317 2019). Therefore, further experimental evidence would be necessary in order to elucidate
318 the role of this genus in mortality events in marine azooxanthellate invertebrates.

319 Overall, our results show that together with *Vibrio* coral pathogens, other specific
320 indicators should be used to assess marine invertebrates' heat stress and *Ruegeria* is likely
321 a good candidate.

322 In the sponge *S. fasciculatus*, the changes related to health status were less evident
323 compared to Anthozoan species. Sponge microbiome has been described to be dominated
324 by *Proteobacteria* with *Chloroflexi*, *Cyanobacteria* and *Crenarchaeota* occasionally
325 reaching high relative abundances (Thomas et al., 2016). In the current study, almost the
326 same phyla were present in healthy and necrosed sponge microbiomes, which were
327 dominated by *Proteobacteria* and *Poribacteria*, although some differences could be
328 detected at genus level. *Acidobacteria* Subgroup 10 became dominant in necrosed
329 samples (Figure 4B and Table S6) compared to healthy samples. This increase was due
330 to OTU8 (Table S5), which was closely related to an uncultured *Acidobacteria* clone
331 (FJ269280.1) isolated from the sponge *Xestospongia testudinaria* in Indonesia (Montalvo
332 and Hill, 2011). This finding suggests that this OTU could be shared by taxonomic and
333 geographically distant sponge hosts and could be a generalist symbiont within the core
334 sponge microbiome. An increase of *Synechococcus* genus was also detected in necrosed
335 sponges (Figure 4B and Table S6). However, one of the main OTUs responsible for the
336 differences among healthy and necrosed samples in *S. fasciculatus* was OTU57 related to
337 *Candidatus Synechococcus spongiarum* (Slaby and Hentschel, 2017). This OTU was a
338 characteristic of healthy samples, where it was 3-fold more abundant than in necrosed
339 ones (Table S5). *Candidatus S. spongiarum* was previously reported as one of the most

340 common symbionts in this sponge and its abundance was related to the increase of
341 seawater temperature (Erwin et al., 2012). Similar shifts in *S. fasciculatus* associated
342 bacteria community composition have been reported previously during the 2010 summer
343 disease episode in the Mediterranean Sea with higher abundances of *Acidobacteria* and
344 lower abundances of *Candidatus S. spongiarum* (Blanquer et al., 2016), although the
345 diseases signs observed by these authors (small white spots) were different to the tissue
346 necrosis reported in the current study. Thus, it seems that an increase of *Acidobacteria*
347 Subgroup 10 and a decrease of *S. spongiarum* could be indicators of heat stress in *S.*
348 *fasciculatus*, although more studies are necessary in order to understand their role in the
349 sponge diseases development.

350

351 **Core microbiome of Anthozoans in the Mediterranean Sea**

352 Previous studies (van de Water et al., 2017; 2018a) have demonstrated that the
353 microbiome of Mediterranean Anthozoans largely depends on their location and could
354 influence their hosts' adaptation to new environmental conditions. Therefore, in order to
355 ascertain the core microbiome associated with each of the three Anthozoans studied here
356 (*C. caespitosa*, *O. patagonica* and *L. sarmentosa*) throughout the Mediterranean Sea, we
357 have analyzed a total of 98 16S rRNA libraries from different Mediterranean locations
358 (Table S1), including a total of 143,281 OTUs. The analysis indicated that the core
359 microbiome of these three Anthozoans species in the studied area was composed of 43
360 OTUs in *C. caespitosa*, 45 in *O. patagonica* and 43 in *L. sarmentosa*, while this core
361 microbiome throughout the Mediterranean Sea was reduced to 4 OTUs in *O. patagonica*
362 (3 *Ruegeria* OTUs and 1 *Pseudovibrio* OTU), 4 in *C. caespitosa* (2 *Ruegeria* OTUs, 1
363 *Pseudovibrio* OTU and 1 *Vibrio owensii* OTU) and 9 in *L. sarmentosa* (5
364 *Endozoicomonas* OTUs, 1 *Ruegeria* OTU, 1 BD1-7 clade, 1 *Granulosicoccus* and 1

365 *Winogradskyella*). Thus, as expected, the increase found in the biogeographical area
366 studied implies a decrease of the corresponding core microbiome. The presence of
367 *Pseudovibrio* OTUs in *O. patagonica* and *C. caespitosa* and *Endozocionomonas* OTUs in
368 *L. sarmentosa* core microbiomes confirms that they are stable bacterial symbionts that
369 are less sensitive than other members of the community to the surrounding environment
370 and they could be good indicators of their hosts' health. Furthermore, these results
371 confirm that, although there is a core microbiome of Anthozoans in the Mediterranean
372 Sea, there is also a local core, as previously observed in Mediterranean gorgonians (van
373 de Water et al., 2017). Thus, our findings confirm that the definition of the core
374 microbiome must be associated with the geographical area considered in the analysis.

375 Importantly, only one OTU, highly similar to *Ruegeria* OTU119, composed the core
376 microbiome of these three Anthozoans. As mentioned above, *Ruegeria* OTU119
377 increased its abundance in unhealthy samples of all Anthozoans studied here. Although
378 this genus has been previously related to the spread of coral diseases worldwide (Sekar
379 et al., 2008; Sunagawa et al., 2009; Apprill et al., 2013) its role in coral microbiome is
380 still unclear. Indeed, this genus is not only present in the core microbiome of healthy
381 specimens *O. patagonica*, *C. caespitosa* and *L. sarmentosa* throughout the
382 Mediterranean Sea, is also commonly associated with a large number of other coral
383 species around the world (Huggett and Apprill, 2018; Rothing et al., 2020), even in
384 larvae forms (Sharp et al., 2012; Zhou et al., 2017). It has been recently demonstrated
385 that species belonging to this genus associated with corals show antibacterial activity
386 against *Vibrio* coral pathogens (Miura et al., 2019) and provide essential vitamins like
387 cobalamin (Karimi et al., 2019). Taken together, this OTU composed the core
388 microbiome of geographically distant Anthozoans in the Mediterranean Sea and its

389 abundance increased in necrotic tissues of their host under heat stress, highlighting the
390 importance of this genus in marine invertebrate health during MHWs.

391 **Figure legends**

392 **Figure 1.** Upset plot showing the relationship of OTUs identified in all marine
393 invertebrates and seawater samples analyzed in this study. A) Graph of the OTUs
394 average (X axis) in each sample (Y axis). B) Intersection of sets of OTUs in each
395 sample. The number of OTUs in each set appears above the column, while the sample
396 shared are indicated in the graph below the column by a point, with the samples on the
397 left. Intersection in red represent OTUs shared by all hosts and seawater samples and
398 intersection in blue OTUs shared by the three Antozoan.

399 **Figure 2** (A) Shannon diversity index and (B) OTU richness obtained for host-
400 associated and surrounding water microbiomes based on 16S rRNA gene diversity.

401 **Figure 3.** Principle coordinate analysis (PcoA) 2D plot based on microbial communities
402 associated with *Cladocora caespitosa*, *Oculina patagonica*, *Leptogorgia sarmentosa*
403 and *Sarcotragus fasciculatus* tissues clustered using coordinated analysis of the
404 weighed UniFrac distance matrix. A) The x- and y-axes are indicated by the first and
405 second coordinates, respectively, and the values in parentheses show the percentages of
406 the community variation explained. B) The x- and y-axes are indicated by the first and
407 third coordinates, respectively, and the values in parentheses show the percentages of
408 the community variation explained

409 **Figure 4.** Overview of the composition of the microbiome composition and microbial
410 community changes related to tissue necrosis signs associated with *Cladocora*
411 *caespitosa*, *Oculina patagonica*, *Leptogorgia sarmentosa* and *Sarcotragus fasciculatus*

412 at (A) class and (B) genus level. For full taxonomic information refer to Supplementary
413 Supplementary Data 1.

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425 **Ethics declarations**

426 **Conflict of Interest**

427 On behalf of all authors, the corresponding author states that there is no conflict of
428 interest.

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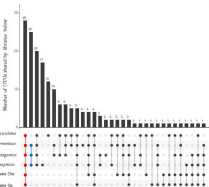
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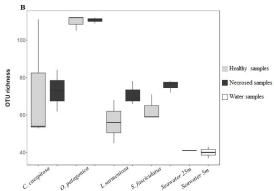
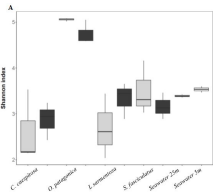
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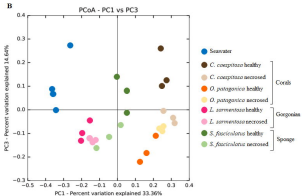
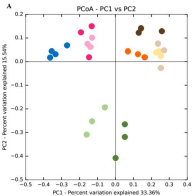
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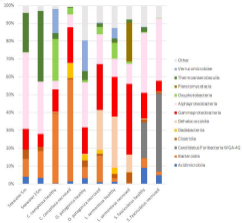


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