The effect of sewage sludge containing microplastics on growth and fruit development of tomato plants

Ricardo Hernández-Arenas, Ana Beltrán-Sanahuja, Paula Navarro-Quirant, Carlos Sanz-Lazaro

PII: S0269-7491(20)36468-X

DOI: https://doi.org/10.1016/j.envpol.2020.115779

Reference: ENPO 115779

To appear in: Environmental Pollution

Received Date: 4 June 2020

Revised Date: 28 September 2020

Accepted Date: 3 October 2020

Please cite this article as: Hernández-Arenas, R., Beltrán-Sanahuja, A., Navarro-Quirant, P., Sanz-Lazaro, C., The effect of sewage sludge containing microplastics on growth and fruit development of tomato plants, *Environmental Pollution*, https://doi.org/10.1016/j.envpol.2020.115779.

This is a PDF file of an article that has undergone enhancements after acceptance, such as the addition of a cover page and metadata, and formatting for readability, but it is not yet the definitive version of record. This version will undergo additional copyediting, typesetting and review before it is published in its final form, but we are providing this version to give early visibility of the article. Please note that, during the production process, errors may be discovered which could affect the content, and all legal disclaimers that apply to the journal pertain.

© 2020 Elsevier Ltd. All rights reserved.







	Journal Pre-proof
1	The effect of sewage sludge containing microplastics on growth and fruit
2	development of tomato plants
3	Ricardo Hernández-Arenas ¹ ; Ana Beltrán-Sanahuja ^{2*} ; Paula Navarro-Quirant ² ,
4	Carlos Sanz-Lazaro ^{1,3} .
5	¹ Department of Ecology, University of Alicante, PO Box 99, E-03080 Alicante, Spain.
6	² Analytical Chemistry, Nutrition & Food Sciences Department, University of
7	Alicante, PO Box 99, E-03080 Alicante, Spain.
8	³ Multidisciplinary Institute for Environmental Studies (MIES), University of
9	Alicante, PO Box 99, E-03080 Alicante, Spain.
10	
11	
12	
13	
14	*All correspondence should be addressed to this author. Tel: +34 965909645.
15	Fax: +34 965903527. E-mail: <u>ana.beltran@ua.es</u>
16	
17	
18	
19	
20	
21	
22	
23	
24	

25 ABSTRACT

26 Microplastics (MPs) are becoming an environmental growing concern, being the sewage sludge applied to agriculture fields one of the most important inputs to the 27 28 environment. To date, there is no standardized protocol for their extraction and changes 29 in vegetative growth and fruit maturation on cultivated plants induced by sludge 30 containing MPs have not been studied yet. Sewage sludge from three different 31 wastewater treatment plants located in Murcia, Spain, were studied. First, the 32 microplastic concentration was estimated and, then, the effects of the sewage sludge in 33 the development of tomato plants and fruit production was analyzed. The measured 34 parameters in tomato plants were both, biomass and length, for shoot and root part, as 35 well as, stem diameter and tomato production. The present work has developed and 36 validated a protocol for the extraction and quantification of MPs comprising several 37 shapes, materials and sizes from samples of sewage sludges, which offers a good compromise for the extraction of different types of microplastic. The protocol used for 38 39 MPs extraction had a recovery efficiency of 80 ± 3 % (mean \pm SE) and used 40 bicarbonate, to maximize MPs extraction. The mean abundance of MPs in the studied sewage sludge samples was $30,940 \pm 8,589$ particles kg⁻¹ dry weight. Soils with sludge 41 42 containing MPs fostered the growth of tomato plants, while delaying and diminished 43 fruit production. However, other factors or their interactions with MPs could have 44 influenced the outcomes. Further studies are necessary to corroborate these findings and 45 explain the mechanisms of possible effects of MPs on plants.

46 Capsule: In this work, the effects of sewage sludges containing microplastics on growth
47 and fruit development in tomato plants have been evaluated.

48

49 **KEYWORDS:** Tomato plants, sewage sludge, microplastics, biomass, fruit crop

50 1. Introduction

51 Microplastics (MPs) are an environmental growing concern, with an annual input to 52 the environment of 11 million tons (Boucher and Friot, 2017). Agriculture fields 53 receiving sludge from sewage treatment plants as fertilizers constitute one of the largest 54 inputs of MPs to the environment (63000-430000 t of MPs per year; (Nizzetto et al., 55 2016). Only in Europe, wastewater treatment plants (WWTPs) produce 13 million tons of dry weight sludge per year and approximately 99% of MPs can be retained in the 56 57 sludge (Magnusson and Norén, 2014). 80% of this sludge is used in agriculture (Mahon 58 et al., 2017). Despite the importance of quantifying the concentrations of MPs in sludge 59 (de Souza et al., 2018; Corradini et al., 2019), to date, there is no standardized protocol for their extraction and quantification (Rocha-Santos and Duarte, 2015; Rochman et al., 60 61 2017).

Different methodologies have been proposed for the extraction of MPs from sludge samples, generally using oxidizing agents and acids to remove part of the organic matter and a density separation of the MPs by means of saline solutions (Lusher et al., 2017; Sujathan et al., 2017; Hurley et al., 2018; Lares et al., 2018; Li et al., 2018). However, one of the drawbacks of using high concentrations of these agents to remove organic matter is the possible degradation of some polymers, (Avio et al., 2015), difficulting their identification by using Raman and FT-IR spectroscopy (Lares et al., 2018).

Additionally, estimates of MPs concentrations must be accompanied by a recovery test of the most abundant types of MPs comprising different sizes. Previous studies quantifying MPs in sludge or in soil samples have generally performed partial recovery tests only using one (Mahon et al., 2017; Sujathan et al., 2017; Li et al., 2018) or two shapes of MPs (Hurley et al., 2018; Corradini et al., 2019) and in some cases a recovery test was not applied (Lusher et al., 2017; van den Berg et al., 2020). Nevertheless,

recovery tests are key in assessing the effectiveness of the extraction methods,
necessary to design standardized procedures that can precisely estimate MPs
concentrations in sewage sludge.

MPs are expected to alter the biophysical properties of the soil, in terms of water 78 79 retention capacity, bulk density, microbial activity and formation of aggregates, which 80 could lead to an alteration in the water cycle in the soil (de Souza Machado et al., 2018; 81 Wan et al., 2019). However, the effect that MPs accumulation in the soil will generate 82 on the vegetation has been scarcely studied. MPs present in sludge from WWTPs used 83 in agriculture can remain in soil for a long-time affecting the vegetative growth (van den 84 Berg et al., 2020). In this sense, it has been reported that MPs can induce changes in 85 both aerial and below-ground parts during vegetative growth in the wheat plant (Qi et al., 2018) while reducing the germination rate and root growth of *Lepidium sativum* as 86 87 the concentration of MPs to which plants were exposed increased (Bosker et al., 2019). Although sewage sludge has been employed to improve soil fertility in agriculture 88 89 (Llorens et al., 2012), to the best of our knowledge, changes induced by sludge 90 containing MPs in vegetative growth and fruit maturation of cultivated plants have not 91 been studied.

The aim of this work is to assess the efficiency of a protocol for MPs extraction and quantification from sewage sludge samples and evaluate the effect of sewage sludge containing MPs on tomato plants once applied in soil. To assess the efficiency of the extraction and quantification protocol, a recovery test was carried out using different types of particles, materials and sizes. At last, the effects of sewage sludge containing MPs on tomato plants were evaluated by means of response variables related to growth and fruit production.

100

101 **2. Materials and Methods**

102 **2.1 WWTPs and sludge samples**

103 Sludge samples were obtained from the municipal WWTPs located in Librilla (W1),

104 Alhama de Murcia (W2) and Totana (W3) (Murcia, Spain) (Table S1).

105 **2.2 Microplastics (MPs) extraction procedure from sludge samples**

106 Three replicates of 20 mL sludge from each WWTP were introduced in 250 mL pre-107 washed glass beakers. Afterwards, 40 mL of solution H₂O₂ (30 wt %) was added to the 108 beakers and left for 2 d to remove most of the organic matter present in the sample. 109 Then, MPs were extracted through density separation by adding 150 mL of a NaCl 110 saturated solution and stirring with a glass rod for 3 minutes. The solution was then allowed to settle overnight. Afterwards, the supernatant was decanted and filtered. By 111 112 adding 150 mL of the saline solution, MPs decantation was facilitated by minimizing 113 resuspension of the settled particles.

114 This process was repeated four times. Since the density of the used saline solution was 1.2 g cm⁻³, some polymers, mainly polyethylene terephthalate (PET; (density 1.29 – 115 1.40 g cm⁻³) and polyvinyl chloride (PVC; density 1.30 - 1.58 g cm⁻³) (Duxbury, 1992) 116 117 may not float in the solution but settle on the bottom of the beaker. To prevent this, in 118 the last repetition of the process, 6 g of sodium bicarbonate were added to the saline 119 solution. Then, the solution was heated for approximately 4 minutes until the 120 temperature reached between 45°C and 50°C and stirred with a glass rod to facilitate the 121 release of CO₂ bubbles that promote the flotation of MPs to the top of the saline 122 solution (per. obs.). The supernatant was filtered through glass fiber filters with a pore 123 size of 1.2 µm (Frisenette ApS, Knebel, Denmark).

124 Filters were left in petri dishes and the remaining organic matter was removed by 125 submerging them with approximately 20 mL of 1:1 (v/v) H_2O_2 (30 wt. %) + H_2SO_4 (96 126 wt. %) (Merck, Darmstadt, Germany) solution and left overnight. Since the H₂O₂:H₂SO₄ 127 solution produces a very exothermic reaction, it was left to cool down before being used 128 to prevent possible damage to MPs. The used solution was chosen after several tests 129 using the following proportions of H_2O_2 (30 wt. %) and H_2SO_4 (96 wt. %): 1:1, 1:2 and 130 1:3 aiming to find the best compromise between organic matter removal and avoidance 131 of PET microfibers disintegration, which are the most sensitive component to be 132 disintegrated out of the main plastic debris found in the sludge (per. obs.). The best 133 results were obtained by using the 1:1 ratio in which fibers were not visually affected. 134 Higher proportions of H₂SO₄ lead to total or partial disintegration of the fibers.

135 After the exposition of filters to the acid solution, the liquid content in the petri 136 dishes was filtered. Both petri dishes and filters were rinsed off several times with distilled water to ensure that no particles were retained. Then, particles in the filters 137 138 were identified by Raman spectroscopy and quantified. Since some particles were 139 retained on the walls of the bottom part of the funnel of the filtering equipment, after 140 each filtration, the funnel was separated from the main structure, washed with distilled 141 water and the liquid collected was filtered again in the same filtration unit to avoid MPs 142 losses.

143

144 **2.3 Recovery test**

A recovery test was performed to estimate the extraction efficiency and reproducibility of the protocol used in this study. The recovery test was carried out using commercially available plastic materials with different shapes and sizes (Table S2) representative of the MPs found in sludge. The MPs were spiked into 20 mL of

sludge from W3. The particles used in the recovery test had specific colors and shapes to facilitate their differentiation from particles already present in the sludge. In addition, it was previously checked that the treatments applied to remove the organic matter did not alter the particles used in the recovery test.

153 Fragments were composed of high density polyethylene (HDPE) and polypropylene 154 (PP) from a bleach bottle and a cap of a container, respectively. For films, HDPE and 155 low density polyethylene (LDPE) were used, from different plastic bags. The sizes of 156 fragments and films were: 0.4-0.6 mm, 0.8-1.0 mm, 1.9-2.1 mm and 2.4-2.6 mm. PET 157 microfibers were obtained from two T-shirts with a length of ca. 2 mm. Microbeads 158 were composed of butylene/ethylene copolymer (B/E) and they were obtained from a 159 shower gel presenting a size between 0.6 and 0.8 mm. Fragments, films and microfibers 160 were cut using a binocular magnifying glass (LFZ Dual, OPTECH, Ontario, Canada), a 161 graduated scale and a scalpel to produce the desired sizes. To test the repeatability, a 162 total of 10 particles of each size and type of material were added to 20 mL of sludge, 163 summing a total of 80 fragments, 80 films, 20 PET microfibers and 10 B/E microbeads. 164 The recovery test was performed six times to test its reproducibility, indicating the 165 calculated variation in the extraction efficiency when the measurements are reproduced.

166 The recovery efficiency percentage was calculated as follows:

167 $E(\%) = (P / Pr) \cdot 100$

Where E stands for the Efficiency, P refers to the number of particles added and Pris the number of particles recovered.

170 2.4 Characterization of MPs by Raman Spectroscopy

171 MPs were characterized by Raman spectroscopy following the procedure indicated in

172 the Supplementary Data. Spectra of the suspected MPs were compared to the Raman

173 spectra of reference materials (Figure S1), (Boerio et al., 1976; Andreassen, 1999;

174 Long, 2004; Kida et al., 2016) (Table S3).

175 **2.5. Microplastics count**

The MPs present in each filter were counted using a binocular magnifying glass (LFZ Dual, OPTECH, Ontario, Canada) with the aid of a transparent circular template divided in eight sections that was placed at the top of the Petri dish.

179 **2.6. Quality control**

During the whole process of MPs extraction, blank filters were exposed to the air as quality controls to evaluate possible cross-contamination (Woodall et al., 2015). Particles found in these filters were analysed through Raman spectroscopy as explained previously. See Supplementary Data for a more detailed explanation.

184

185 **2.7. Plant material and growth conditions**

186 The tomato plant (Lycopersicon esculentum Mill.) was chosen since it is widely cultivated and shows low or no toxicity due to over fertilization with macronutrients 187 (Sainju et al., 2003). One-month tomato plants were transplanted outdoors on the 14th of 188 189 March of 2019 individually into a 3.3 L pot made of terracotta. The dimension of each 190 pot was 19 cm high, diameter of 20.7 cm at the top and 12.7 cm at the bottom. Five 191 soils, control (C), manure control (MC), and soils containing sludge of each WWTP 192 (W1, W2 and W3), were created using 50:50:0:0, 45:45:10:0, and 45:45:0:10 193 proportions of peat moss (COMPO SANA ® UNIVERSAL, Münster, Germany), silica sand (grain size range: 0.4-0.8 mm), manure (FEMABÓN ®, Castellón, Spain) and 194 195 sludge, respectively.

196 Control soil was used as the reference substrate, while manure control soil was used 197 as a fertilizer-type amendment to soil aimed to simulate nutrient addition to the soil but 198 with an insignificant MPs concentration. To ensure that there was no nutrient deficiency 199 due to the different element composition of manure and sewage sludge in all soils, each 200 plant was fertilized monthly with 10 g of a solid slow-release fertilizer (NPK 12-8-18) 201 (Bluefficient Platinnum®, Salamanca, Spain), aiming to prevent changes in the 202 development of tomatoes due to a specific element deficiency while avoiding fertilizer 203 toxicity. Characteristics of the employed peat, manure, and fertilizer used are shown in 204 Table S4. Each type of soil was replicated in seven pots, each one containing a tomato plant, so seven tomato plants were cultivated in each type of soil which made a total of 205 206 35 tomato plants.

207 The pots containing the plants were placed outdoors in an area of the campus of the 208 University of Alicante that had direct radiation from the sun during the whole day. A 209 wood stick was tied to each plant to prevent them from bending and a blue protective 210 mesh was placed in each pot to avoid plants from animal attacks. The experiment lasted 109 d. Each pot was watered with 500 mL of tap water. Watering frequency was once 211 212 every two days until June. Then, it was increased to twice per day until the experiment finished on the 1st of July. The position of the pots was changed every month to avoid 213 214 possible effects due to placement.

215

216 **2.8 Measurement of pH and soil salinity**

Every month, pH and soil salinity were measured to monitor nutrient availability and water absorption capacity of the tomato plants as it is described in Supplementary Data.

221 **2.9. Collection of tomatoes**

The fruits of all tomato plants began to be harvested after three and a half months of cultivation and the fruit selection was carried out according to their maturation status with the purpose to collect tomatoes in a similar maturation stage, using as a criterion the color of tomato fruits (Figure S2).

226 **2.10. Measurements of tomato plant growth parameters**

227 Once the experiment was concluded, tomato plants were taken out of the pots and 228 roots were cleaned in a cube with water to remove soil particles attached. Shoot length 229 of tomato plants was measured as the maximum length of the main stem, whereas root 230 length was determined as the maximum length of the longest root. Stem diameter was 231 measured with a digital vernier caliper (Mitutoyo CD-6 C5, Neuss, Germany). Then, 232 tomato plants were cut into the shoot and root parts and dried during 72 h at 70°C to 233 obtain the dry weight of each part.

234

235 2.11. Statistical analysis

236 The experiment consisted of a one factor (type of soil) design with five types of 237 soils: C, MC, W1, W2 and W3. Each treatment was replicated three times. Significant 238 differences among the response variables measured in the different types of soils were 239 analyzed through a one-way analysis of variance (ANOVA). Before ANOVA, the 240 normality was tested through Kolmogorov-Smirnov's test and heterogeneity of variance 241 of the data was tested using Cochran's test. If the data did not meet these assumptions, 242 transformations were applied, and assumptions were checked again. When significant 243 differences were found in the ANOVA main test, the *post-hoc* Tukey's test was used to 244 find which treatments significantly differed (Table S5). Statistical analyses were performed with the R statistical software (v. 3.6.0; R Core Team, 2019). All statistical 245

246 tests were conducted with a significance level of $\alpha = 0.05$. Data were reported as mean 247 \pm standard error (SE).

248

250

252

- 249 **3. Results and discussion**
- 251 **3.1 Method extraction efficiency**

The mean average extraction efficiency for the total of particles spiked in the sludge 253 254 samples was $87 \pm 2\%$ (Figure 1). The range of particle size with the highest extraction 255 efficiency was 2.4-2.6 mm (91 \pm 5%) whereas the lowest extraction efficiency was 256 obtained in particles with size in the range 1.9-2.1 mm (87 ± 3 %). Fragments and films 257 were extracted with an average of 91 ± 2 % and 86 ± 3 %, respectively. Microfibers had 258 a mean extraction efficiency value of $83 \pm 3\%$. Spiked microbeads showed $90 \pm 6\%$ of 259 extraction efficiency. As regarding the type of material PP, HDPE, LDPE and PET were 260 extracted with an efficiency of $91 \pm 2\%$, $89 \pm 2\%$, $88 \pm 3\%$ and $83 \pm 3\%$ respectively.

261 Figure 1

262 The obtained percentage values of extraction efficiency are, in general, comparable 263 with previous studies (Mahon et al., 2017; Sujathan et al., 2017; Hurley et al., 2018; Li 264 et al., 2018; Corradini et al., 2019). Only in some cases are slightly lower, which could 265 be due to the wide range of sizes of particles used in this study, generally reaching 266 lower sizes than other studies (Huerley et al., 2018). In addition, the recovery test 267 carried out in this study is comprehensively complete compared to previous ones where 268 only a few types of particles, materials and sizes were used (Mahon et al., 2017; 269 Sujathan et al., 2017; Hurley et al., 2018; Li et al., 2018;). Specifically, in this work the 270 recovery test uses microfibers, which is generally the main component of sludge despite 271 not being a particle type commonly used in recovery tests (Mahon et al., 2017; Sujathan 272 et al., 2017; Li et al., 2018). Our results indicate that this extraction method is capable

of efficiently recover PP, LDPE, HDPE and PET microfibers, which are the most commonly used plastics (PlasticsEurope, 2019) and the most common types of MPs found in sludge (Mahon et al., 2017; Li et al., 2018; Corradini et al., 2019). Thus, this study demonstrates that the present extraction method is suitable for the extraction of MPs in sewage sludge, while being inexpensive and safer than others that use other salts, such as ZnCl and NaI.

279 280

3.2 Characterization of microplastics in sludge samples

The concentration of MPs was $17,870 \pm 2,174$; $27,821 \pm 1,357$ and $47,130 \pm 3,002$ particles kg⁻¹ dry weight of sludge in W1, W2 and W3, respectively (Figure 2). The sizes of the MPs were in the range of 0.31-2.11 mm (Figure S2).

284 **Figure 2**

W3 showed a significantly higher concentration of MPs in comparison with W1 and
W2. Concerning MPs shape, microfibers were the most abundant particle followed by
fragments, films and microbeads.

Figure 3

289 Microfibers represented the 93 \pm 2 %, 89 \pm 5 % and 97 \pm 1% of the total MPs in 290 W1, W2 and W3, respectively. All microfibers analyzed were made of PET, so it is 291 suspected that they were synthetic fibers used for acrylic garment, since microfibers in 292 sludge mainly come from cloth washing (Browne et al., 2011; Folkö, 2015; Åström, 293 2016). Additionally, microfibers, since they are made mostly of PET, their density, 294 higher than water and along with their high surface:volume ratio, they are prone to 295 precipitate in the sludge during wastewater processing in the WWTP. This explains that 296 PET microfibers were the most abundant MPs in sludge as previously reported (Mahon 297 et al., 2017; Li et al., 2018; Corradini et al., 2019).

302 The lesser abundance of fragments and films in sludge in comparison to fibers sides 303 with previous studies (Mahon et al., 2017; Li et al., 2018; Corradini et al., 2019). This 304 may be due to PE, has lower density than water and it tends to float, and it is less likely 305 that their fate is the sludge. Additionally, because of its shape, fragments and films, for 306 a specific size, are more likely to be retained than microfibers in the pretreatment filters 307 of WWTPs before entering in the degreaser-desander channel. In fact, W2, which was 308 the WWTP with the largest mesh size pretreatment filter (70 mm) of all the WWTPs, 309 had the greatest percentage of fragments $(9 \pm 4\%)$ and films $(2 \pm 1\%)$ compared to the 310 percentage of fragments found in W1 (6 \pm 2 %) and W3 (2 \pm 1 %), which had a mesh 311 size of 2 and 3 mm, respectively. These findings seem a plausible explanation of the 312 low presence of fragments and films in sludge. Microbeads only constituted 0.4 ± 0.4 %, 0.4 ± 0.4 % and 0.2 ± 0.2 % of the total MPs in W1, W2 and W3, respectively as 313 314 previously reported(Mahon et al., 2017; Li et al., 2018).

As regards the extraction process in the sludge samples, out of the total particles extracted, 61 ± 4 %, 19 ± 4 %, 13 ± 2 % and 7 ± 2 % of the particles were recovered in the first, second, third and fourth extraction step, respectively. In the last extraction step, bubbles produced by the bicarbonate kept microfibers on the surface of the solution used for the separation of MPs through difference of density. Thus, the use of bicarbonate for MPs extraction could be advisable to maximize the process, especially in matrices with high microfibers content.

322 The average concentration of MPs in the sludge amongst the three WWTPs was $30,940 \pm 8,589$ particles kg⁻¹ dry weight, being higher than previous studies(Corradini et 323 324 al., 2019); (van den Berg et al., 2020); (Mahon et al., 2017).

325

3.3. Measurements of tomato plant parameters

327 After 109 d growth, significant differences among soil treatments were found in 328 biomass production, but not in the morphological variables studied (Figure 4).

329 Figure 4

326

330 Shoot biomass was significantly higher in W2 (79.2 \pm 7.3 g) than in C (35.8 \pm 4 g), 331 MC soil (41.4 \pm 3.4 g) and W3 (53.4 \pm 5.1 g). Significant differences were also found

between W1 (60.9 \pm 7.2 g) and C samples. For root biomass, W2 (67.8 \pm 7.3 g) was 332

333 significantly higher than C (23.4 \pm 5.1 g), MC (36 \pm 10.4 g) and W3 (32.4 \pm 5.1 g).

334 Total biomass was significantly higher in W2 (146.9 \pm 11.3 g) than in C (59.1 \pm 8.7 g),

335 MC (77.4 \pm 11.8 g) and W3 (85.8 \pm 11.2 g) soils. Significant differences were found

336 between W1 (60.9 \pm 7.2g) and C samples. The values of stem diameter, height and root

337 length ranged between 11.7 \pm 1.5 and 10.3 \pm 1.1 mm, 74.0 \pm 5.4 and 61.9 \pm 15.4 cm and 338 34.2 ± 6.0 and 27.4 ± 6.8 cm, respectively.

339 The number of tomatoes harvested on each tomato plant did not significantly vary as 340 regards the type of soil. MC soil had the highest tomatoes production with 5 ± 1 341 tomatoes while W3 produced less than 2 fruits. During the experiment, the number of 342 mature tomatoes produced in plants grown in C and MC soil was significantly higher 343 $(1.4 \pm 0.3 \text{ and } 1 \pm 0.4 \text{ tomatoes, respectively})$ than the ones grown in soils containing 344 sludge with MPs (W1; 0.13 ± 0.13 ; W2: 0.14 ± 0.14 ; W3: 0) (Figure 5).

345 Figure 5

At the end of the experiment, MC had the highest percentage of plants (75%) that produced tomatoes, whereas in C or soils containing sludge from W1, W2 and W3 this percentage was lower being 50, 37.5, 25 and 25%, respectively.

349 The mean values of pH obtained for W1 (6.4 ± 0.1) and W3 (6.5 ± 0.1) were 350 considered slightly acidic, whereas W2 (6.6 \pm 0.4), MC (7.2 \pm 0.1) and C (6.8 \pm 0.1) were 351 considered neutral (Juárez et al., 2004), which limits nutrient deficiency problems of 352 nutrient availability related to highly acidic or basic soils (Sainju et al. 2003). The mean 353 values of electrical conductivity obtained from C (3.3 \pm 0.2 dS/m), W1 (2.9 \pm 0.2 354 dS/m), W2 (2.8 \pm 0.1 dS/m) and W3 (2.7 \pm 0.1 dS/m) were considered very slightly saline (2-4 dS/m) whereas MC soil was considered slightly saline (5.3 \pm 0.2 dS/m) 355 356 (Schoeneberger et al., 2002). The sludge addition to soils generally results in an 357 increase of their salinity, but in case of saline soils, it can reduce their salinity (Pomares 358 et al. 1998). Peat moss, that was used as the base of the soil component in the present experiment, can retain a large amount of cations, which could be responsible of 359 360 increasing the levels of salinity. The variety of tomato plants used, Raf, have a threshold 361 of tolerance to salinity of 6 dS/m (Nelson, 2011). Thus, the levels of salinity in the 362 different types of soil is not expected to negatively affect the growth and yield of tomato 363 plants.

Nutrient levels in the different sludge used were in some cases notably different. For example, N was $7.6\pm0.9\%$ in W2, while in W1 and W3 was 1.9 ± 0.2 and 5.4 ± 0.6 , respectively (Table S1). High contents of N concentration can promote biomass at the expense of fruit yield (Sainju et al. 2003). Our results indicate a notable shoot and root growth of tomato plants in soil W2, which was significantly different to the plants grown in soil W3, but not in W1, which had the lowest N content (Fig. 4). Nevertheless, fruit production was not affected in plants grown in any of the soils treated with sludge(Fig. 5).

Phosphorous and K values were also different among the used sludges. Therefore, in our experiment a solid slow-release fertilizer containing N, P and K was applied to all the treatments aimed to prevent nutrient limitation. Other macronutrients, such as Mg and Ca, despite the content being different among sludges, the levels were high enough that no deficiency was expected. Nutrient toxicity due to the above-mentioned nutrients is not expected in tomato plants, apart from the commented issue with N (Sainju et al. 2003).

Our results suggest contrasting effects of sludge addition in soils since the biomass 379 380 production was higher in plants cultured with sludge than the ones cultured in the C or 381 MC soils. However, tomato plants cultured in soils treated with sludge from W3, which 382 had the highest concentration of MPs, had the lowest biomass and tomato production, 383 while it did not produce any mature tomato during the experiment. Plant biomass can be influenced by many factors, such as soil humidity, soil and air temperature, 384 385 photoperiod, solar radiation, precipitations, genotype, etc. One of the most important 386 factors influencing biomass is soil nutrient availability since both nutrient deficiency 387 and toxicity negatively affect total biomass and fruit production (Msilini et al., 2009; 388 Karim et al., 2012). The solid slow-release fertilizer aimed to prevent nutrient 389 limitation. The shoot:root biomass ratio of cultured plants did not show significant 390 differences among plants cultured in the different soils, suggesting non-relevant 391 differences in nutrient availability among soils (Ericsson, 1995). Water was homogeneously provided, while it was ensured that all the plants were similarly 392 393 exposed to light.

394 The effects of MPs in soil have been scarcely studied, but are expected to be highly 395 influenced by the shape (Rilling et al., 2019) and size (Bosker et al., 2019), being 396 microfibers among the type of MPs that is expected to have the strongest effects (de 397 Souza Machado et al., 2019). In the present experiment, differences in plant biomass 398 among soils with sludge were noticeable. Root mass is expected to increase in the 399 presence of microfibers since they can lower soil bulk density (de Souza Machado et al., 400 2018) promoting increased root growth due to reduced penetration resistance for plant 401 roots and improved soil aeration (de Souza Machado et al., 2019). These differences 402 could also directly and indirectly affect soil structure through the modification of soil 403 aggregation (de Souza Machado et al., 2018). Our results side with previous findings 404 showing the largest biomass in the root, as well as, in shoot in soils with sludge, 405 especially in W2 (de Souza Machado et al., 2018; de Souza Machado et al., 2019). Very 406 high concentrations of MPs, as in W3, can have the opposite effect, reducing root 407 biomass production (Jiang et al., 2019).

408 Our results indicate a lower crop in the tomato plants grown under soil treated with 409 sludge at the time of the end of the experiment. This could be because of microfibers in 410 increasing C:N ratio as a consequence of the modification of nutrient availability 411 derived from alterations in water dynamics reported in onion plants (de Souza Machado 412 et al., 2019). Nutrient alterations can lead to plant stress, lowering crop production (Li 413 et al., 2009). Since the experiment did not last until the end of the crop season, we could 414 not know if the tomato plants suffered a reduction in crop or only a delay in crop 415 production. Nevertheless, a delay in the crop production is expected to lower the overall 416 crop production that is possible to be harvested (Thomison et al., 2011).

417 When interpreting the results of this study, possible variables that were not 418 controlled, such as nutrient level, could influence them and need to be taken into

419 account. Despite the solid slow-release fertilizer provided to all the treatments aimed to 420 prevent nutrient limitation, nutrient concentration was not equal in the different 421 treatments. Further experiments, aiming to have a more similar nutrient concentration 422 among treatments would be desirable.

423 Additionally, sewage sludge is a complex matrix that does not only contain MPs, 424 but also metals, pathogens and organic toxicants. The levels of metal concentration 425 found in sludge samples were below the ones established in the European Council 426 Directive 86/278/EEC (Table S1). Sludge samples also contained bacteria such as 427 Escherichia coli and Salmonella sp., that inhabit human intestinal tracts (Table S1) and 428 so are not necessarily pathogens, but certain strains can cause pathogenicity. Thus, 429 nutrients, metals, bacteria and other variables, as well as the interactions of these 430 variables with MPs could have influenced the results of this study. Because the above 431 commented variables are not controlled, the experimental design of our study using 432 different sludges does not allow us to demonstrate a direct cause-effect of MPs. 433 However, this work allows us to test the real effects of sewage sludge containing MPs. 434 Complementary studies would be needed to increase the insight on the possible effects 435 of MPs from sewage on agriculture.

This study suggests, that despite the large number of MPs reported in sludge (Lusher et al., 2017; Mahon et al., 2017; Li et al., 2018; Corradini et al., 2019; van den Berg et al. 2020), these estimates could underestimate real concentrations or indicate that the number of MPs is growing. Under this scenario of high amounts of MPs exported through sewage sludge to the environment (De Souza Machado et al. 2018), it is urgent that policy makers define maximum levels of MPs in sludge used for agriculture as it has already been done with other toxic compounds, such as metals (86/278/EEC).

To the best of our knowledge, no study has tested the effects of sludge with MPs in the yield of agriculture crops. Our results show that sewage sludge with large concentrations of MPs can notably affect tomato plants, which could be due to a greater or lower extent to MPs. However, other possible factors may have been responsible or can produce additive or interactive effects with MPs. Thus, further work is necessary to corroborate these findings and to explain the mechanisms of possible effects of MPs on plants, such as tomato.

450 Author Contributions:

- 451 Conceptualization, C.S., A.B.,; methodology R.H., C.S., A.B., P.N.; formal
- 452 analysis, R.H., A.B., C.S., ; investigation, R.H., P.N., A.B., C.S., ; resources, A.B., C.S.;

453 data curation, A.B., C.S., R.H.; writing-original draft preparation, R.H., A.B., C.S.,;

454 writing—review and editing, C.S., A.B., ; supervision, A.B., C.S. All authors have

455 read and agreed to the published version of the manuscript.

456 Funding

457 C. S. was funded by the University of Alicante (Ref. UATALENTO 17-11).

458 **Declaration of competing interest**

459 The authors declare that there are no conflicts of interest

460 ACKNOWLEDGEMENTS

461 The authors would like to express our gratitude to the company FACSA S.A
462 (Castellón, Spain) for kindly providing us with the WWTP samples from Totana,
463 Alhama de Murcia and Librilla (Murcia, Spain) municipal WWTPs. The authors would
464 like to thank Natalia Sánchez García and Susana Carrión Jaén for helping during the
465 experiment.

	Journal 110-proof
467	
468	References
469	
470	Andreassen, E., 1999. Infrared and Raman spectroscopy of polypropylene. pp. 320–328.
471	https://doi.org/10.1007/978-94-011-4421-6_46
472	Åström, L., 2016. Shedding of synthetic microfibers from textiles.
473	Avio, C.G., Gorbi, S., Regoli, F., 2015. Experimental development of a new protocol
474	for extraction and characterization of microplastics in fish tissues: First
475	observations in commercial species from Adriatic Sea. Marine Environmental
476	Research 111, 18–26. https://doi.org/10.1016/j.marenvres.2015.06.014
477	Boerio, F.J., Bahl, S.K., McGraw, G.E., 1976. Vibrational analysis of polyethylene
478	terephthalate and its deuterated derivatives. Journal of Polymer Science: Polymer
479	Physics Edition 14, 1029–1046. https://doi.org/10.1002/pol.1976.180140607
480	Bosker, T., Bouwman, L.J., Brun, N.R., Behrens, P., Vijver, M.G., 2019. Microplastics
481	accumulate on pores in seed capsule and delay germination and root growth of the
482	terrestrial vascular plant Lepidium sativum. Chemosphere 226, 774–781.
483	https://doi.org/10.1016/j.chemosphere.2019.03.163
484	Boucher, J., Friot, D., 2017. Primary microplastics in the oceans: A global evaluation of
485	sources. IUCN International Union for Conservation of Nature.
486	https://doi.org/10.2305/IUCN.CH.2017.01.en
487	Browne, M.A., Crump, P., Niven, S.J., Teuten, E., Tonkin, A., Galloway, T.,
488	Thompson, R., 2011. Accumulation of microplastic on shorelines worldwide:
489	Sources and sinks. Environmental Science & Technology 45, 9175–9179.
490	https://doi.org/10.1021/es201811s
491	Corradini, F., Meza, P., Eguiluz, R., Casado, F., Huerta-Lwanga, E., Geissen, V., 2019.
492	Evidence of microplastic accumulation in agricultural soils from sewage sludge
493	disposal. Science of the Total Environment 671, 411–420.
494	https://doi.org/10.1016/j.scitotenv.2019.03.368
495	de Souza Machado, A.A., Lau, C.W., Kloas, W., Bergmann, J., Bachelier, J.B., Faltin,
496	E., Becker, R., Görlich, A.S., Rillig, M.C., 2019. Microplastics can change soil
497	properties and affect plant performance. Environmental Science & Technology 53,
498	6044–6052. https://doi.org/10.1021/acs.est.9b01339
499	de Souza Machado, A.A., Lau, C.W., Till, J., Kloas, W., Lehmann, A., Becker, R.,
500	Rillig, M.C., 2018. Impacts of microplastics on the soil biophysical environment.
501	Environmental Science & Technology 52, 9656–9665.
502	https://doi.org/10.1021/acs.est.8b02212
503	Duxbury, M.A., 1992. Plastic pellets in the aquatic environment sources and
504	recommendations.
505	Effectson, 1. (1995). Growin and shoot: foot faile of seedings in felation to nutrient
507	availability. In Nutrient uptake and cycling in forest ecosystems (pp. 205-214).
507	Springer, Dordrecht.
500	FOIKO, A., 2015. Quantification and characterization of fibers enflued from common
510	Synthetic indentials during washing. Hurley P.P. Lusher, A.L. Olsen, M. Nizzette, L. 2018, Validation of a method for
510	extracting microplastics from complex organic_rich_environmental matrices
512	Environmental Science and Technology 52, 7400, 7417
512	https://doi org/10/1021/acs est 8b01517
51 <i>5</i>	Jiang X Chen H Liao Y Ye 7 Li M Klobučar G 2010 Ecotovicity and
515	genotoxicity of polystyrene microplastics on higher plant Vicia faba.

516 Environmental Pollution 250, 831-838. 517 https://doi.org/10.1016/j.envpol.2019.04.055 518 Juárez, M., Sánchez, A., Jordá, J.D., Sánchez, J.J., 2004. Diagnóstico del potencial 519 nutritivo del suelo. Universidad de Alicante, Alicante. 520 Karim, MR.R., Zhang, Y.-Q., Tian, D., Chen, F.-J., Zhang, F.-S., Zou, C.-Q., 2012. 521 Genotypic differences in zinc efficiency of Chinese maize evaluated in a pot 522 experiment. Journal of the Science of Food and Agriculture 92, 2552–2559. 523 https://doi.org/10.1002/jsfa.5672 524 Kida, T., Hiejima, Y., Nitta, K., 2016. Raman spectroscopic study of high-density 525 polyethylene during tensile deformation. International Journal of Experimental Spectroscopic Techniques 1, 1–6. https://doi.org/10.35840/2631-505X/8501 526 527 Lares, M., Ncibi, M.C., Sillanpää, Markus, Sillanpää, Mika, 2018. Occurrence, 528 identification and removal of microplastic particles and fibers in conventional 529 activated sludge process and advanced MBR technology. Water Research 133, 530 236-246. https://doi.org/10.1016/j.watres.2018.01.049 Li, S. X., Wang, Z. H., Malhi, S. S., Li, S. Q., Gao, Y. J., & Tian, X. H. (2009). Nutrient 531 532 and water management effects on crop production, and nutrient and water use 533 efficiency in dryland areas of China. Advances in agronomy, 102, 223-265. 534 https://doi.org/10.1016/S0065-2113(09)01007-4 535 Li, X., Chen, L., Mei, Q., Dong, B., Dai, X., Ding, G., Zeng, E.Y., 2018. Microplastics 536 in sewage sludge from the wastewater treatment plants in China. Water Research 537 142, 75-85. https://doi.org/10.1016/j.watres.2018.05.034 538 Llorens, E., Gallardo, A., García-Agustín, P., Lapeña, L., Molina, M.J., 2012. Response 539 of tomato crops (Solanum lycopersicum 'Montecarlo') to sewage sludge-based 540 compost fertilization. Acta Horticulturae 123–130. 541 https://doi.org/10.17660/ActaHortic.2012.933.13 542 Long, D.A., 2004. Infrared and Raman characteristic group frequencies. Tables and 543 charts. George Socrates John Wiley and Sons, Ltd, Chichester, Third Edition, 544 2001. Journal of Raman Spectroscopy 35, 905–905. 545 https://doi.org/10.1002/jrs.1238 546 Lusher, A.L., Hurley, R., Vogelsang, C., Nizzetto, L., Olsen, M., 2017. Mapping 547 microplastics in sludge. 548 Magnusson, K., Norén, F., 2014. Screening of microplastic particles in and down-549 stream a wastewater treatment plant. 550 Mahon, A.M., O'Connell, B., Healy, M.G., O'Connor, I., Officer, R., Nash, R., 551 Morrison, L., 2017. Microplastics in sewage sludge: Effects of treatment. 552 Environmental Science and Technology 51, 810–818. 553 https://doi.org/10.1021/acs.est.6b04048 Msilini, N., Attia, H., Bouraoui, N., M'rah, S., Ksouri, R., Lachaâl, M., Ouerghi, Z., 554 2009. Responses of Arabidopsis thaliana to bicarbonate-induced iron deficiency. 555 556 Acta Physiologiae Plantarum 31, 849–853. https://doi.org/10.1007/s11738-009-557 0318-z 558 Nelson, W., 2011. Influencia de la salinidad y la relación de Calcio/Potasio sobre el 559 crecimiento y desarrollo del tomate cv.Raf. 560 Nizzetto, L., Futter, M., Langaas, S., 2016. Are agricultural soils dumps for 561 microplastics of urban origin? Environmental Science and Technology. 562 https://doi.org/10.1021/acs.est.6b04140 563 Pomares, R., Estela, M., Tarazona, F., 1998. Efecto de diferentes enmiendas orgánicas 564 en las propiedades del suelo de un huerto de cítricos. Agrochimica, 42(1-2), 41-49.

Journal Pre-proof

565	Qi, Y., Yang, X., Pelaez, A.M., Huerta Lwanga, E., Beriot, N., Gertsen, H., Garbeva, P.,
566	Geissen, V., 2018. Macro- and micro- plastics in soil-plant system: Effects of
567	plastic mulch film residues on wheat (Triticum aestivum) growth. Science of the
568	Total Environment 645, 1048–1056.
569	https://doi.org/10.1016/j.scitotenv.2018.07.229
570	Rillig, M.C., Lehmann, A., Ryo, M., Bergmann, J., 2019. Shaping up: Toward
571	considering the shape and form of pollutants. Environmental Science &
572	Technology 53, 7925–7926, https://doi.org/10.1021/acs.est.9b03520
573	Rocha-Santos, T., Duarte, A.C., 2015. A critical overview of the analytical approaches
574	to the occurrence, the fate and the behavior of microplastics in the environment.
575	TrAC Trends in Analytical Chemistry 65, 47–53.
576	https://doi.org/10.1016/i.trac.2014.10.011
577	Rochman, C.M., Regan, F., Thompson, R.C., 2017. On the harmonization of methods
578	for measuring the occurrence, fate and effects of microplastics. Analytical Methods
579	9. 1324–1325. https://doi.org/10.1039/C7AY90014G
580	Sainiu, U. M. Dris, R. Singh, B. 2003. Mineral nutrition of tomato. Food. Agriculture
581	and Environment 1(2) 176-183
582	Schoeneberger, P.J., Wysocki, D.A., Benham, E.C., Broderson, W.D., 2002, Field Book
583	for Describing and Sampling Soils Version 2.0
584	Sujathan S Kniggendorf AK Kumar A Roth B Rosenwinkel KH Nogueira
585	R 2017 Heat and bleach: A cost-efficient method for extracting microplastics
586	from return activated sludge. Archives of Environmental Contamination and
587	Toxicology 73 641–648 https://doi.org/10.1007/s00244-017-0415-8
588	Thomison PR Mullen RW Linns PE Doerge T Gever A B 2011 Corn
589	response to harvest date as affected by plant population and hybrid Agronomy
590	Journal 103 1765–1772 https://doi.org/10.2134/agroni2011.0147
591	van den Berg P Huerta-I wanga E Corradini E Geissen V 2020 Sewage sludge
592	application as a vehicle for microplastics in eastern Spanish agricultural soils
593	Environmental Pollution 261 https://doi.org/10.1016/i.envpol.2020.114198
594	Wan Y Wu C Xue O Hui X 2019 Effects of plastic contamination on water
595	evaporation and desiccation cracking in soil. Science of The Total Environment
596	654 576–582 https://doi.org/10.1016/i.scitoteny.2018.11.123
597	Woodall LC Gwinnett C Packer M Thompson RC Robinson L F Paterson
598	G L L 2015 Using a forensic science approach to minimize environmental
599	contamination and to identify microfibres in marine sediments. Marine Pollution
600	Bulletin 95 40–46 https://doi.org/10.1016/j.marpolbul.2015.04.044
601	Yang Qin M (2014 May) Global fibres overview Synthetic Fibres Raw Materials
602	Committee Meeting at APIC
603	
604	
605	Wehlink
606	
607	ESAMUR 2019 Use of sludge from wastewater treatment in agriculture in the Region
608	of Murcia https://www.esamur.com/lodos-edar (accessed 26 May 2020)
609	or marola. https://www.esumar.com/rodos/edur (accessed 20 may 2020).
610	
611	
612	

613 FIGURES

- 614 Fig. 1. Percentage of particles extracted in the recovery test (mean \pm SE; n=2) 615 concerning size, type of polymer and particle.
- 616 Fig. 2. Number of microplastics (mean \pm SE; n=3) found in sludge samples from the
- 617 wastewater treatment plants in the cities of Librilla (W1), Alhama de Murcia (W2) and
- 618 Totana (W3) in Spain. Different letters on top of the bars represent statistically
- 619 significant differences (Tukey-HSD test, p <0.05).
- Fig. 3. Percentage of types of polymers (A) and particles (B) (mean \pm SE; n=3) found in
- 621 the cities of Librilla (W1), Alhama de Murcia (W2) and Totana (W3) in Spain.
- Fig. 4. Morphological and biomass parameters (mean \pm SE; n=7) after 109 d of growth..
- 623 Different letters on top of the bars represent statistically significant differences (Tukey-
- 624 HSD test, p < 0.05). Control (C), Manure Control (MC), Librilla (W1), Alhama de
- 625 Murcia (W2), Totana (W3).
- 626 Fig. 5. Number of tomatoes (A) and number of mature tomatoes produced (B) (mean \pm
- 627 SE).. Different letters on top of the bars represent statistically significant differences
- 628 (Tukey-HSD test, p < 0.05). Control (C), Manure Control (MC), Librilla (W1), Alhama
- 629 de Murcia (W2), Totana (W3).

- 631
- 632
- 633
- 634
- 635
- 636
- 637







641 Figure 2

















654 SUPPLEMENTARY DATA

655 2.4. Characterization of MPs by Raman Spectroscopy

MPs were characterized by using a Raman spectrometer NRS-5100 (Jasco, Madrid, Spain) equipped with LMU-20X-UVB lens. The laser excitation frequency and intensity used was 784.79 nm and 11.8 mW, respectively. Raman spectra were recorded with a charge-coupled device camera (UV-NIR range, 1024×255 pixels) electrically cooled to -70°C. Raman spectra were obtained between 162 and 1886 cm⁻¹ with a spectral resolution of 2.47 cm⁻¹.

662

663 **2.6. Quality Control**

A total of 29 fibers were found in 3 blank filters. The mean concentration of MPs $(9.7 \pm 0.3 \text{ particles})$ in blank filters were subtracted to the mean value of the extracted MPs.

To test for modifications in the Raman spectra of the samples due to the chemical agents employed to remove the organic matter, a preliminary trial was performed with all the types of polymers used in the recovery test. The Raman spectra of all the types of polymers were obtained before and after performing the extraction protocol performed to the sewage sludge samples. No relevant changes in the intensity of the characteristic bands (Figure S3) were found in comparison with the spectra of the same materials before applying the extraction protocol.

To minimize contamination during the lab procedure all the containers were covered with aluminum foil, lab instrumentals were washed before use and were made of glass, the windows of the laboratory were closed, and 100% cotton coats were worn.

	1.0			
01100		10.1		
Uulli				

677 The Raman analysis of the fiberglass, the material the filter was made of, did not 678 emit fluorescence that could interfere with the analysis of the samples, before and after 679 applying the treatment to remove organic matter.

2.8. Measurement of pH and soil salinity

To determine pH values, air-dried soil samples from the pots were sieved through a 2 mm mesh to keep all primary soil particles and remove roots and bigger soil aggregates. Then, 40 mL of sieved soil was added to a glass beaker and slowly wetted with distilled water (sensu Juárez et al., 2004). After that, soil was stirred for 20 seconds with a glass rod and left 60 minutes before the pH was measured (pH Basic 20, Crison Instruments S.A., Barcelona, Spain). To monitor soil salinity, a container was placed under each tomato plant to collect 50 mL of the water lixiviated after watering the plants. Then, the electrical conductivity of the collected water samples was measured by using an electrical conductivity meter (Crison Basic 30, Barcelona, Spain).

	Journal Pre-proof
702	
703	
704	Table S1. WWTPs and sludge samples characteristics.
705	Table S2. Characteristics of MPs used in the recovery test. Number of particles shown
706	are referenced to one recovery test.
707	Table S3. PE, PET and PP Raman bands utilized for the identification of the
708	microplastics.
709	Table S4. Label information of the products utilized in the composition of the soil
710	treatments.
711	Table S5. Results of the ANOVA for the SL (shoot length), RL (root length), D
712	(diameter), SB (shoot biomass), RB (root biomass), RB:SB (root:shoot biomass ratio),
713	NT (number of tomatoes), NMT (number of mature tomatoes) and MPA (microplastics
714	abundance). The factors were WWTPs (wastewater treatment plants) and TOS (type of
715	soil).
716	Figure S1. Raman spectrum of a fiber and its similarity to a reference material.
717	Figure S2. Photographs of some microplastics found in the sewage sludge samples.
718 719	Figure S3. Tomato appearance and color at harvest time.
720	Figure S4. Raman spectra of a PP fragment (A), LDPE film (B) and PET fiber (C)
721	before (grey line) and after (black line) the reagents used for removing the organic

- 722 matter.

	Journal Pre-proof
727	
728	
729	Table S1. Characteristics of the wastewater treatment plants (WWTP) and sludge. W1, W2 and W3,
730	correspond to the three WWTPs located in Librilla, Alhama de Murcia and Totana, respectively in
731	Murcia,Spain.

	W1	W2	W3
WWTP			
Population equivalent	4,312	17,293	27,628
Type of treatment	Secondary and tertiary	Secondary and tertiary	Secondary and tertiary
Water sources	Domestic	Domestic and industrial	Domestic and industrial
Treatment capacity (m3/year)	227,815	994,596	1,684,947
Sludge production (kg dry sludge /year)	55,632	317,280	695,592
Sludge			
Dry matter (%)	18±2	13±1	18±2
Total organic matter (%)	77±6	84±7	66±5
N (%)	1.9±0.2	7.6±0.9	5.4±0.6
$P(P_2O_5) (mg kg^{-1})$	54,204±10,841	36,846±7,369	37,716±7,543
K (K ₂ O) (mg kg ⁻¹)	9,220±1,660	3,484±627	6,570±1,183
/Ig (MgO) (mg kg ⁻¹)	15,036±2,707	10,384±1,863	15,692±2,825

	Ca (CaO) (mg kg ⁻¹)	31,906±6,381	31,668±6,334	70,714±14,143
732				
733				
734	Continuation Table S1			

	W1	W2	W3
Sludge			L
Cu (mg kg ⁻¹)	189 <u>±</u> 28	186±28	449±67
Cd (mg kg ⁻¹)	<2.0	<2.0	<2.0
Cr (mg kg ⁻¹)	27±5	21±4	32±6
Hg (mg kg ⁻¹)	0.23±0.08	0.24±0.08	0.27±0.09
Ni (mg kg ⁻¹)	20±3	16±2	19±3
Pb (mg kg ⁻¹)	18±3	25±5	40±7
Fe (mg kg ⁻¹)	4689±750	3850±616	8223±1316
$Zn (mg kg^{-1})$	511±87	608±103	1599±272
<i>Escherichia coli</i> (CFU/g; confidence interval)	140000±88000- 220000	140000±88000- 220000	17000±11000- 27000
Salmonella sp.(presence/25g)	absence	presence	presence

	Journal Pre-proof
741	
742	
743	
744	Table S2. Characteristics of MPs used in the recovery test. Number of particles shown
745	are referenced to one recovery test.
746	

Polymer type	Number of particles	Particle shape	Color	Source	Size (mm)
HDPE	40	Fragment	Blue	Bleach bottle	0.4-0.6, 0.8-1, 1.9 2.1, 2.4-2.6
HDPE	40	Film	Black	Plastic bag	0.4-0.6, 0.8-1, 1.9-2.1, 2.4-2.6
LDPE	40	Film	Blue	Plastic package	0.4-0.6, 0.8-1, 1.9-2.1, 2.4-2.6
РР	40	Fragment	Red	Bottle tap	0.4-0.6, 0.8-1, 1.9-2.1, 2.4-2.6
PET	10	Fiber	Blue	100% polyester T- shirt	2
PET	10	Fiber	Yellow	100% polyester T- shirt	2

	B/E	10	10 Microbead		Palmolive Thermal Spa Mineral Massage ®	0.6 - 0.8
747 748 749 750 751 752	Table S3. Exa	amples of t	he Raman bands	utilized fo	or the identification o	f PE, PET and PP.

Table S3. Examples of the Raman bands utilized for the identification of PE, PET and PP.

Polymer	Wavenumber (cm ⁻¹)	Assignment	Reference		
PE	1063	Anti-symmetric stretching (C-C)	Kida et al., (2016)		
	1080	Stretching (C-C)			
	1130	Symmetric stretching (C-C)			
	1298	Twisting (C-C)			
	1440	Bending (CH ₂)			
	1460	Bending (CH ₂)			
PET	1290	Stretching C(O)-C	Boerio et al., (1976)		
	1414	CCH bending and OCH bending			
	1615	Ring mode 8a			
	1726	Stretching C=O			

754 755

	Journal Pre-proof
756	
757	
758	
759	
760	
761	
762	
763	Continuation Table S3

Polymer	Wavenumber (cm ⁻¹)	Assignment	Reference
PP	809	Rocking CH ₂ , Stretching CC _b , Stretching C-CH ₃	Andreassen (1999)
	841	Rocking CH ₂ , Stretching CC _b , Stretching C-CH ₃ , rocking CH ₃	
	900	Rocking CH ₃ , rocking CH2, bending CH	
	941	Rocking CH ₃ , stretching CC _b ,	
	973	Rocking CH ₃ , stretching CC _b ,	
	998	Rocking CH ₃ , bending CH, wagging CH ₂	
	1040	Stretching C-CH ₃ , Stretching CC _b , bending CH	
	1152	Stretching CC _b , stretching C- CH ₃ , bending CH, Rocking CH ₃	
	1219	Twisting CH ₂ , bending CH, stretching CC _b	

	Journal Pre-proof	
1330	Bending CH, twisting CH ₂	
1360	Symmetric bending CH ₃ , bending CH	
1458	Asymmetric bending CH ₃ , bending CH ₂	

764 b: backbone



Table S4. Label information of the products utilized for the composition of the soil

767 treatments.

Product	Label information
COMPO SANA ® UNIVERSAL	Composition: black peat, Perlita, Agrosil ® and lime; pH (CaCl ₂): 5,0 – 6,5; Salt content (KCl) g/l: <3; Content in subscription (soluble nutrients): 200 – 400 mg/l Nitrogen (N), 200 – 500 mg/l Phosphate (P ₂ O ₅), 300 – 500 mg/l Potassium oxide (K2O)
Manure FEMABÓN ®	Composition: herbaceous peat and compost manure; organic material (dry weight): 48%; bulk density: 610 g/l; electrical conductivity: 3 dS/m pH: 7.9
BLUEFFICIENT PLATINNUM®	12% Nitrogen; 7,5% (N) ammoniacal; 4.5% (N) ureic; 8% P ₂ O ₅ ; 18% K ₂ O; 2% MgO 17% SO ₃ ; 0.1% Mn; 0.1% Zn 3%; pH 6.0 inhibitor
200	

799	
800	

Table S5. Results of the ANOVA for the SL (shoot length), RL (root length), D (diameter), SB (shoot biomass), RB (root biomass), RB:SB (root:shoot biomass ratio), NT (number of tomatoes), NMT (number of mature tomatoes) and MPA (microplastics abundance). The factors were WWTPs (wastewater treatment plants) and TOS (type of soil).

			SL		RL		D		SB				
Source of variation	d.f.	MS	F	Р	MS	F	Р	MS	F	Р	MS	F	Р
TOS	4	245.4	1.574	>0.2	51.85	1.507	>0.2	2.261	0.843	>0.5	2020	8.351	<0.001
Residual	30	155.9			34.40			2.681			242		
Total	34												
Cochran's C test			C = 0.30043,		(C = 0.39329	l,	C	C = 0.49131,			C = 0.3611	.8,
			P > 0.05			P > 0.05			P < 0.05			P > 0.05	
Transformation			None			None			None			None	
Tukey's test												W2 > C	>0.0001
												W2 > MC	>0.001
												W2 > W3	>0.03
												W1 > C	>0.03
			RB			RB:SB			NT			NMT	
Source of variation	d.f.	MS	F	Р	MS	F	Р	MS	F	Р	MS	F	Р
TOS	4	2065.5	6.041	<0.01	0.1124	1.105	>0.3	8.896	0.513	>0.7	0.7198	4.755	<0.01
Residual	30	341.9			0.1017			17.343			0.1514		
Total	34												
Cochran's C test			C = 0.37135,		(C = 0.69274		C	C = 0.30208,			C = 0.4362	.7,
			P > 0.05			P < 0.05			P > 0.05			P > 0.05	
Transformation			None			None			None			logarithm	ic
Tukey's test			W2 > C	>0.0008								C > W1	>0.01
			W2 > MC	>0.03								C > W2	>0.02
			W2 > W3	>0.009								C > W3	>0.005
			MPA										
Source of variation	d.f.	MS	F	Р									
WWTPs	2	663975790	34.97	<0.001									
Residual	6	18987671											
Total	8												
Cochran's C test			C = 0.47455,										
			P > 0.05										
Transformation			None										
Tukey's test			W3 > W1	> 0.03									
			W3 > W2	> 0.0004									





Fig. S2. Photographs of some microplastics found in the sewage sludge samples.



865	
866	
867	
868	Fig. S3. Tomato appearance and color at harvest time.
869	
870	
871	
872	
873	
874	
875	
876	
877	
878	
879	
880	
881	
882	
883	
884	
885	
886	
88/	
000	
009 900	
890 801	
807	
893	
894	
895	
896	
897	



Fig. S4. Raman spectra of a PP fragment (A), LDPE film (B) and PET fiber (C) before
(grey line) and after (black line) the exposition to the reagents used for removing the
organic matter.

HIGHLIGHTS

An extraction protocol for MPs from sewage sludge has been validated.

Up to $31,000 \pm 8,600$ particles kg⁻¹ dry weight in sewage sludge were estimated.

Sewage sludge containing MPs fostered the growth of tomato plants.

Sewage sludge containing MPs delayed and diminished fruit production.

building

Author Contributions:

Conceptualization, C.S., A.B.,; methodology R.H., C.S., A.B., P.N.; formal analysis, R.H., A.B., C.S., ; investigation, R.H., P.N., A.B., C.S., ; resources, A.B., C.S.; data curation, A.B., C.S., R.H.; writing—original draft preparation, R.H., A.B., C.S.,; writing—review and editing, C.S., A.B., ; supervision, A.B., C.S. All authors have read and agreed to the published version of the manuscript.

., A.B., C.S.

Declaration of interests

 \boxtimes The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests: